

Trade Growth, Production Fragmentation, and China's Environment

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Abstract

In recent years, China has experienced both rapidly growing trade and serious environmental degradation. The large literature on trade and environment lends some credence to the idea that these are causally related: trade growth for a relatively poor country is thought to shift the composition of industrial output towards dirtier products, aggravating environmental damage. However, much of China's trade growth is attributable to the international fragmentation of production and the growing dominance of trade in parts and components—fragments. This kind of trade could lead to “cleaner” trade if fragmented production occurs in cleaner goods or China specializes in cleaner stages of production within these goods. Using official Chinese environmental data on air and water pollution from the State Environmental Protection Agency and highly disaggregated trade data from China Customs, we present evidence that the pollution intensity of Chinese exports fell dramatically between 1995 and 2004. We then explore the possibility that trade fragmentation and foreign investment have played a role. Using the framework provided by Copeland and Taylor (1994, 2003), we develop a reduced form model of the pollution intensity of trade, incorporating standard determinants of a country's production mix, such as factor proportions, income per capita, and trade policy. We explicitly incorporate the degree to which Chinese exports are fragmented, building on the work of Feenstra and Hanson (1996). We then use this model to test the effect of increased fragmentation on the time trends we observe in the pollution intensity of trade. The evidence supports the view that increased foreign investment and production fragmentation have contributed positively to the decline in the pollution intensity of China's trade, as has accession to the WTO. Growth in China's per capita real income is also associated with the trend toward cleaner trade.

I. Introduction

In recent years, China has been notable for its rapidly growing trade and its serious environmental degradation. China's trade with the world has risen dramatically between 1995 and 2005. In current dollars, the value of China's exports plus imports rose from \$280.9 billion in 1995 to \$1422.1 billion in 2005--a growth of about 406%. While major improvements have been made in water and air quality over the same period, China's State Environmental Protection Agency (SEPA) stated that "[t]he conflict between environment and development is becoming ever more prominent. Relative shortage of resources, a fragile ecological environment and insufficient environmental capacity are becoming critical problems hindering China's development." (*SEPA, 2006*).

Some of the large literature on trade and environment lends credence to the idea that trade growth and environmental degradation are causally related. The environmental Kuznets curve literature suggests that low income countries have relatively lenient environmental standards, and hence a comparative advantage in pollution-intensive goods.¹ As a low-income country grows, environmental damage increases due to increased scale of production, and a composition of output biased towards "dirty goods." However, higher incomes also generate pressure for more stringent environmental regulations. Since tighter regulations raise the cost of polluting and give producers incentives to find cleaner production techniques, this tends to reduce environmental damage.² For low income countries, the scale and composition effects are thought to outweigh the technique effect, implying that the net effect of growth is detrimental to the environment. Since trade growth raises incomes, it, too, contributes to these scale, composition and technique effects. Yet empirical evidence on the net effect of trade and environmental damage is mixed, with at least some studies (Dean, 2002; Antweiler, et al., 2001) finding evidence that the technique effect may be stronger than previously

¹ The evidence on the existence of an environmental Kuznets curve is mixed, and highly dependent upon time period, countries evaluated, and pollutants examined. Thus, there is no way to verify whether or not China is to the left or right of the turning point in the "inverted U." For a recent survey of this evidence, see Copeland and Taylor (2004). For surveys covering the broader literature on trade and environment, see Dean (2001) and Copeland and Taylor (2004).

² In addition, some would argue that increased FDI would imply greater environmental degradation, as firms in pollution-intensive industries may move to avoid more stringent environmental regulations at home. See Dean, Lovely, Wang (2006) for review of evidence and counterargument.

thought, leading to a net beneficial impact of trade growth on the environment.

China's integration with the world economy may not fit this conventional picture. Much of China's trade growth is attributable to the international fragmentation of production--the splitting of production processes into discrete sequential activities (fragments) which take place in different countries³ (Chen, et al., 2004; Ping, 2005; Dean, Fung, Wang, 2006). China's trade statistics explicitly designate "processing imports" as imports of intermediate inputs to be used to produce products solely for export, and "processing exports" as those exports which use these imported inputs.⁴ This trade alone accounts for about 56% of the growth in China's exports and 41% of the growth in China's imports between 1995 and 2005. In addition, a large part of this trade is attributable to foreign-invested enterprises (FIEs).⁵ In 2005, about 84% of China's processing exports and imports were carried out by FIEs.

Trade arising from international production fragmentation could be cleaner than conventional trade. If highly fragmented industries (such as computers and other high-tech products) and the particular fragments within these products that China produces are relatively clean, then China's output and trade would shift toward cleaner goods as these activities expand. In addition, if the FIEs who carry out much of this trade in fragments produce using greener technologies than those used by domestic producers in China, production techniques within fragmented industries would become cleaner over time. In this way, both the composition and technique effects of trade growth may be favorable to China's environment.

This chapter explores these relationships using new evidence on the pollution content of Chinese trade. We first present evidence on the growth of trade and industrial emissions in China. Using official Chinese environmental data on air and water pollution from the State Environmental Protection Agency, we find that industrial emissions of primary pollutants have slowed or fallen over the last decade while trade has grown. Across most industrial sectors, the pollution intensity of production has also fallen. We then explore

³ See Arndt and Kierzkowski (2001) for discussion of the causes of fragmentation.

⁴ Chinese trade statistics record two types of processing imports and exports: processing and assembly (where the foreigner retains ownership of imported inputs), and processing with imported inputs (where the importer acquires ownership of imported inputs).

⁵ Chinese trade statistics record several types of FIEs: fully-funded enterprises (i.e., wholly-owned subsidiaries of foreign companies), equity joint ventures, and contractual joint ventures.

trends in the pollution intensity of Chinese trade. Building on highly disaggregated trade data from China Customs, we report new evidence that the pollution intensity of Chinese exports has fallen dramatically from 1995 to 2004. We use a counterfactual exercise to show that this decrease in the pollution intensity of trade is due partly to a shift in the composition of trade toward cleaner goods, but also to a shift in production technique toward cleaner processes.

Finally, we explore the possibility that production fragmentation and processing trade may have played a role in making China's trade cleaner. Building on the framework provided by Copeland and Taylor (1994, 2003), we develop a reduced form model of the pollution intensity of trade, incorporating standard determinants of a country's production mix, such as factor proportions, income per capita, and trade policy. We then incorporate a fragmented export sector, building upon the work of Feenstra and Hanson (1996). The impact of fragmentation on the pollution intensity of China's exports and imports is tested econometrically using the data on four pollutants over a ten year period. We find evidence consistent with the view that the increased importance of processing trade and the extensive presence of foreign invested enterprises have both contributed to reducing the pollution content of China's trade.

II. Trends in Chinese Industrial Emissions and Manufacturing Trade

A. Aggregate trends

In this chapter our interest is in the relationship between China's trade and China's environment, rather than the global environment. Hence, we focus on the primary pollutants which China uses to evaluate the condition of its own environment, rather than the greenhouse gases associated with global climate change. In the 10th Five Year Plan (2001-2005), the Chinese government stated explicit goals for the reduction of its water pollution, as measured by Chemical Oxygen Demand (COD) and its air pollution, as measured by SO₂ and total suspended particles (especially those generated by smoke and dust) (OECD, 2005), COD measures the mass concentration of oxygen consumed by chemical breakdown of organic and inorganic matter in

water.⁶ COD emissions account for the majority of industrial water pollution levies collected in China during this period. While emissions of other water pollutants are recorded in more recent years, they are generally positively correlated with COD. Industrial SO₂ emissions include the sulfur dioxide emitted from fuel burning and from the production processes on the premises of an enterprise. The concentration of SO₂ is a commonly used measure of air quality. Industrial smoke (or soot) emissions includes smoke emitted from fuel burning on the premises of an enterprise. Industrial dust emissions refers to the volume of dust suspended in the air, and emitted by an enterprise's production processes.⁷ Smoke and dust are major contributors to particulate matter (PM¹⁰), which is a commonly used measure of air quality.

Figure 1 shows the trends in China's overall merchandise trade (billions of \$US (2000)) and industrial emissions (billions of kilos) from 1995-2005. Trade data are Chinese official data obtained from China Customs. Industrial emissions data are from the *Chinese Environmental Yearbook* and *China Statistical Yearbook on Environment* (various issues). In Chinese official statistics, the industrial sector includes Mining, Manufacturing and Production and Distribution of Electricity, Gas and Water.⁸ Emissions data prior to 1998 were recorded only for industrial enterprises at the "county level and above." After the "Investigation on Sources of Township Industrial Pollution," published in 1997, it was found that township and village industrial enterprises (TVIEs) were accounting for a larger and larger percentage of emissions. Therefore, emissions data include these enterprises from 1998 onwards. In Figures 1 and 2 we have been able to include TVIE emissions for 1995 and for 1997. But the TVIE data are unavailable for 1996, so we treat 1996 as missing (indicated by the dashed lines).

The most remarkable trend in figure 1 is the dramatic and rapid increase in the value of China's merchandise exports plus imports over the period. At the same time, industrial emissions are generally

⁶ China Statistical Yearbook on Environment, 2006, p. 207.

⁷ *China Statistical Yearbook on Environment*, 2006, p. 208. This does not include indirect generation of dust emitted by using energy generated from power plants.

⁸ Changes in Chinese industrial emissions should be fairly representative of air pollution emissions, since industry accounts for at least 80% of SO₂, smoke, and dust emissions throughout the period. Chinese industrial water pollution emissions accounted for 60% of COD emissions at the start of the period. With emissions from households and services growing in importance, industry's share fell to only 40% by the end of the period. However, total COD emissions fell simultaneously by 20%.

falling. While figure 1 shows a small increase in SO₂ emissions over the period, emissions of COD, soot and dust show a slow but significant decline. This decline in industrial emissions is confirmed in the ten-year environmental review issued by SEPA (2006), and is also noted by the WTO (2006) and the OECD (2005). Figure 2 shows an index of trade and industrial emissions levels, with 1995 as the base year. By 2005, trade had increased nearly 300% in real terms over its 1995 value. Meanwhile annual industrial emissions of COD, smoke, and dust had declined to 56%, 46% and 40%, respectively, of their levels in 1995. In contrast, industrial SO₂ emissions rose after 1999, and were 17.5% above 1995 levels by 2005.

B. Trends in the composition of China's trade

To understand what is driving these aggregate trends, we first examine the trends in the composition of China's trade. Because data on emissions by industrial sectors are readily available, but data for agricultural or service sectors are not, we limit our analysis to manufacturing trade. In 2005, manufacturing trade accounted for 97% of Chinese exports and 83% of Chinese imports. Table 1 shows the shares of exports and imports in 1995 and 2004, by 2-digit ISIC sectors in manufacturing. The Chinese trade data were aggregated to HS (6-digit) and then converted to ISIC Revision 3 using the official Chinese concordance.

Even at this rather aggregated level, table 1 reveals some dramatic shifts in the sectoral composition of Chinese trade over this time period. In 1995, textiles and apparel accounted for the largest shares of Chinese exports to the world. These shares fell by about a third by 2004, while the export share of office and computing machinery grew by a factor of five, and that of communications equipment more than doubled. The largest shares of Chinese imports in 1995 were attributable to textiles and machinery. These shares fell by about 70 % and 40%, respectively, by 2004, while import shares in office and computing machinery and in communications equipment more than doubled.

The sectoral shift in the composition of China's trade is interesting not only because it is dramatic, but because the same sectors have shown increases in both export and import shares. This suggests that much growth has taken place in sectors where production is internationally fragmented, resulting in two-way trade in "fragments" at varying stages of production. One rough indicator of the degree to which industries are internationally fragmented is the share of processing exports (imports) in each sector's total trade. Textile and

apparel exports had substantial shares of processing exports across sectors in 1995, which fell somewhat by 2004. In contrast, office equipment and computing and communications equipment had extremely high shares of processing exports in 1995, and these shares remained high in 2004. China's textile imports showed a decline in the share of processing imports, while communications equipment imports showed a significant rise in processing share from 1995 to 2004. However, this kind of pattern does not occur for machinery imports nor for office and computing machinery imports. This evidence suggests that China's exports (and to a lesser extent imports) have become more concentrated in highly fragmented sectors, and that the degree of fragmentation in some of these sectors has grown over time.

C. Trends in industrial pollution intensity

To see the extent to which changes in production technology could be impacting emissions, we measure the pollution intensity of production by industry, from 1995-2004. We compiled data on emissions of the four pollutants at the industry level, as well as current value of output of the sampled enterprises, from the *Chinese Environmental Yearbooks* (Chinese editions). Pollution intensities were then calculated as emissions (in kilos) per thousand yuan output (constant 1995 yuan) for the 30 Chinese 2-digit industries ("divisions") in the Chinese 2002 industrial classification.⁹ These pollution intensities are shown in appendix table A1. The appendix also provides a detailed explanation of these calculations, and the treatment of missing or aggregated data. In Table 2 we present these average water and air pollution intensities (in kilos per thousand yuan output (constant 1995 yuan)), mapped to the ISIC 2-digit sectors, for 1995 and in 2004.¹⁰

⁹We measure pollution intensity as emissions relative to the value of output because the trade data are also measured in terms of value and our main concern is to measure the pollution intensity of the trade bundle. For some analyses of industrial pollution intensity, a measure of emissions per unit of value-added might be preferable. We are unable to express pollution intensity relative to value added because value-added data are not available at a sufficiently disaggregated level. A comparison of the two measures could reveal important, but unknown, differences. Because the emissions data are classified by economic activity, the numerator of these two measures should be similar as they are not affected by changes in the value of purchased intermediates used in the production process. However, the denominators will differ if an increase in purchased intermediates increases the value of output, thereby reducing pollution measured relative to total value but not relative to value added.

¹⁰ The official Chinese concordance maps the Chinese 2002 industrial classification at the 4-digit level ("classes") to ISIC revision 3 at the 4-digit level. Though some ISIC 2-digit sectors correspond to a single Chinese 2-digit "division," most correspond to several Chinese divisions. Thus, the average pollution intensities for the ISIC 2-digit sectors in table 2 generally represent a weighted average of the pollution intensity of multiple divisions. The weights are the share of all Chinese classes mapped to an ISIC 2-digit sector corresponding to each division. While weights corresponding to

Pollution intensities for manufacturing (ISIC 15-36) and for utilities (ISIC 40-41) are included in the table.¹¹ In each year, the three sectors with the highest pollution intensities are shown in bold for each pollutant.¹²

Of the manufacturing industries, the major source of water pollution is production of paper and paper products. A few others--food products and beverages, and wood products--show relatively high water pollution intensities, but these are far below that of the paper sector. Most industries show very low water pollution intensity. With respect to air pollution, non-metallic minerals (which includes cement) is by far the most SO₂-intensive, and among the top three in terms of smoke and dust. The other industries with high air pollution intensities include basic metals and paper (SO₂), paper and wood (smoke) and wood and furniture (dust). But again these industries generally show much lower pollution intensities than non-metallic minerals. Most industries, in fact, show very low air pollution intensities. The utilities as a group are highly polluting. The water utility is second only to paper production in water pollution intensity. The electricity and gas utilities are the dirtiest sectors overall in terms of SO₂ and smoke.¹³

Table 2 also reveals two interesting trends. The first is that across nearly all sectors, the pollution intensity of production has fallen over time. This is true for all four pollutants. Even the water and energy utilities show improvement over the period. Thus, there is some evidence of a shift toward cleaner industrial production techniques in China. The second trend is that China's trade does appear to be shifting toward cleaner sectors over time. Although trade in 1995 was not concentrated in the highest polluting sectors, textiles and leather products were somewhat high in terms of water pollution intensity, and certainly not the lowest in terms of SO₂ and smoke intensity. Though these industries show cleaner production techniques by 2004, they remain significantly more polluting than office and computing machinery and communications equipment. The latter sectors' pollution intensities were low in 1995 and extremely low as of 2004.

Chinese output by class would be ideal, these data are unavailable.

¹¹ ISIC 37 (Recycling activities) is omitted. See appendix for discussion.

¹² Because there are fewer ISIC 2-digit sectors than Chinese divisions, there is some variation between the highest pollution intensities in table 2 and table A1.

¹³ ISIC Revision 3 groups the electricity and fuel gas utilities into ISIC 40, and as a result, the dust intensity for ISIC 40 looks quite low. But fuel gas production and supply has the second highest dust intensity across Chinese divisions.

the 11 years in the sample (1995-2004). The Chinese pollution intensities¹⁴ for each division are mapped to export (import) shares, with the shares calculated at the ISIC 4-digit level.¹⁵ These 4-digit-weighted pollution intensities are then summed to yield an export (import)-weighted average pollution intensity for each year.

Even by 1995, it appears that Chinese trade liberalization had shifted production toward relatively less water-polluting and less air-polluting products (figure 3). Production of Chinese exports emits about 15 kilos of COD per thousand dollars of exports (in constant 2000 dollars). In contrast, had imports been produced by Chinese import-competing industries, they would have generated 33% more COD emissions (20 kilos per thousand dollars of imports) in 1995. This difference diminishes over time, but remains throughout the period. Chinese exports were also less SO₂-intensive, and less smoke-intensive, than Chinese imports from 1995-2004, but the differences were fairly small. Only if pollution intensity is measured with respect to dust emissions, do we find Chinese exports dirtier than imports.

Figure 3 also shows that both exports and imports become steadily cleaner throughout the period.¹⁶ By 2004, the water pollution intensity of exports had fallen by about 87% while that of imports has fallen by 90%, compared to 1996 levels. The drop in air pollution intensity is almost as dramatic, with export (import) SO₂ intensity falling by 76% (79%), smoke intensity by 80% (80%), and dust intensity by 77% (78%).¹⁷

To understand the relative role of composition and technique effects in generating these trends in pollution intensity, we conduct a counterfactual experiment. We recalculate the pollution intensity of both exports and imports at the ISIC 3-digit level, assuming the pollution intensity of sectoral output remained at its 1995 levels. We then sum these to produce new yearly weighted-average export (import) pollution intensities, shown by the dashed lines in figure 3. These dashed lines show the change in pollution intensity of exports (imports) if only the composition of traded products had changed over time. For all four pollutants,

¹⁴ The pollution intensities were first converted to kilos per thousand US dollars (constant 2000 dollars).

¹⁵ Because the Chinese official concordance maps Chinese 2002 4-digit classes to the ISIC at the 4-digit level, we take advantage of this additional detail to construct more detailed trade weights. See note 10.

¹⁶ Ederington, et al. (2004) find that US exports and imports also have become cleaner over time. Levinson (2007) also finds evidence that both US exports and imports have become cleaner over time, and that technique effects appear to account for more of this change than composition effects.

¹⁷ See previous note. The peak in dust emissions intensity is largely due to the inclusion from 1998 onwards of emissions from TVIEs. Because TVIE emissions data are unavailable at the sectoral level, the yearly industrial pollution intensities in 1995-1997 do not include TVIEs.

changes in the composition of trade did imply both cleaner exports and imports. However, it is also clear that these composition effects account for a relatively small proportion of the observed changes in the pollution intensity of trade. This suggests that China's cleaner production techniques have been the most important force behind cleaner trade.

Because table 1 shows a shift in the composition of China's trade toward highly fragmented manufacturing sectors, and because table 2 suggests that these sectors are relatively low polluters, we examine more closely the pollution intensity of processing trade (figure 4). Several features are immediately evident. First, similar to overall exports, processing exports also tend to be cleaner than imports with respect to all pollutants. Second, processing exports and imports show similar downward trends in pollution intensity to overall trade. Third, the counterfactual results suggest that once again, composition effects are responsible for a small share of the decline in pollution intensity over time. However, most notable is the fact that China's processing trade is much cleaner than China's overall trade. The average water pollution intensity of processing exports (imports) is one-third (60%) that of overall exports (imports), even in 1995. Air pollution intensities are also dramatically lower—only 50% or less-- than those for overall trade. This evidence is suggestive that the increase in China's processing trade has implied a composition effect that is favorable toward China's environment. This effect might be further magnified if the firms engaged in processing trade (largely foreign-invested firms) actually produce with cleaner techniques than average firms.

IV. The Role of Fragmentation and FDI in Explaining the Pollution Intensity of Chinese Trade

To explore the role that production fragmentation and foreign investment play in the changes we observe in the pollution intensity of China's trade, we develop a model that embeds China into the global production network. Our model is tailored for the Chinese context in that it recognizes the magnitude of foreign investment and its effects on the composition of trade. The framework we use draws upon the structural model of pollution developed by Copeland and Taylor (1994), and the outsourcing model developed by Feenstra and Hanson (1996). We first consider the supply of pollution to identify the determinants of pollution regulation. Next, we examine the demand for pollution, first considering the

pollution intensity of exports in a simple two-sector model without fragmented production and then adding a fragmented export sector. We use these models to explore the impact of foreign investment and trade liberalization on the pollution content of trade. Our goal is to derive several reduced form models of the determinants of the pollution intensity of Chinese trade, which we then test empirically.

A. Pollution Supply

We follow Copeland and Taylor (2003) in modeling the supply of pollution as the result of government behavior that maximizes the utility of a representative citizen. Specifically, the Chinese government maximizes indirect utility, V , with respect to its choice of pollution tax rate, \mathbf{t} :

$$V = u(R) - \mathbf{g}D. \quad (1)$$

Here, D is the level of environmental damage experienced by the representative citizen and R is real per capita income. The government takes as given world prices, trade policy, and production possibilities. Real income includes income from production and pollution tax revenue, $(\mathbf{t}D)$, which is rebated to each of the L consumers. Income from production is represented by a GNP function giving the maximized value of national income net of taxes as a function of domestic prices, the pollution tax rate, and vector of factor endowments: $G = G(p, \mathbf{t}, v)$. The first-order condition for maximization of (1.1) yields:

$$\mathbf{t} = \frac{\mathbf{g}Lp}{u'(R)}, \quad (2)$$

where the right-hand side gives the marginal damage from pollution.¹⁸ Using equation (2), we express the endogenous pollution tax as $\mathbf{t}(L, p, R)$.

B. Pollution Demand without Production Fragmentation

We begin with the simplest model of production and trade. This model serves as an alternative to a second model, presented below, that explicitly incorporates export processing with imported intermediate

¹⁸ Because we have adopted a specification in which the marginal disutility of pollution is constant, the pollution supply curve is horizontal. See Copeland and Taylor (2003) for further discussion and alternative specifications.

inputs. We consider a two-sector model of a small, open economy. China is endowed with sector-specific capital and effective labor (E), which depends on the human capital of its labor force: $E = A(H)L$. The import-competing sector, M , uses effective labor and sector-specific capital and it serves as numeraire. Each unit of M produced releases one unit of pollution emissions.

The export sector produces Good Y using effective labor and sector-specific capital (K_Y). Effective labor may also be used for abatement of the pollution emissions (D) created in the production process. Following Copeland and Taylor's (2003) form for abatement, we may express the production function for X treating emissions as an input:

$$y = \left[E_Y^{1-b(z)} D_Y^{b(z)} \right]^q K_Y^{1-q}, \quad (3)$$

where $0 < \mathbf{b} < 1$. The relative domestic price of Y is $p = \mathbf{d} p^*$, where $1/\mathbf{d}$ is a measure of trade frictions and p^* is China's terms of trade. We can use (3) to solve for the pollution intensity of export production, which we denote by e :

$$e \equiv \frac{D}{py} = \frac{\mathbf{b}q}{t}. \quad (4)$$

We use (4) to create our first estimating equation for the pollution intensity of Chinese exports. In doing so, we note that the pollution intensity given by (4) depends on the pollution intensity of China's export production, as measured by the term, \mathbf{b} . As Copeland and Taylor (2003) discuss, differences across countries in factor abundance interact with regulatory differences to determine the pattern of trade. These considerations lead to an expression for the pollution intensity of Chinese exports of the form

$$e_Y = e_Y(K, H, L, t) = e_Y(K, H, L, R, p^*, \mathbf{d}). \quad (5)$$

In this expression we have replaced the pollution tax rate with its determinants, based on (2). Thus, the pollution intensity of exports can be estimated as a function of Chinese factor endowments, its real income per capita, its terms of trade and its trade frictions.

If pollution intensity rises with the capital intensity of production, we would expect China's capital-labor ratio to be positively related to the pollution intensity of its exports but negatively related to the

pollution intensity of its imports.¹⁹ Because an increase in real income raises the level of the pollution tax, we expect the pollution intensity of exports to fall as China's real income rises. The terms of trade and trade frictions have ambiguous effects on pollution intensity. Improved terms of trade imply an increase in real GDP and, hence, a higher domestic pollution tax, reducing e , but a higher relative price for exports raises the production value of factors used in abatement, raising e . If this latter consideration dominates, we would expect improved terms of trade and reduced trade frictions to raise the pollution intensity of China's exports.

C. Pollution Demand with Production Fragmentation

As an alternative to the simple two-sector model above, we consider a model with two export sectors. China is treated as a small economy relative to an Advanced trading bloc. The first sector produces "ordinary" exports, those that are produced with domestic inputs, using the production technology given by (1.3). The "processing" sector produces a set of goods that are intermediate inputs for a single final good. This final good is costlessly assembled from a continuum of intermediate inputs, indexed by $z \in [0,1]$. Inputs are produced using effective labor, capital specific to the processing sector, and pollution discharge. Input production technology varies by the amount of labor used relative to the emissions created during production. We adopt a simple functional form for production technology of input z :²⁰

$$x(z) = \left[E(z)^{1-a(z)} D(z)^{a(z)} \right]^q K(z)^{1-q}. \quad (6)$$

. We also restrict $\mathbf{a}(z) \in [\underline{\mathbf{a}}(z), \bar{\mathbf{a}}(z)]$, $0 < \underline{\mathbf{a}} < \bar{\mathbf{a}} < 1$, and $0 < \mathbf{q} < 1$. We assume that ordinary export production is more pollution intensive than processing export production, implying that $\mathbf{b} > \bar{\mathbf{a}}$.

Intermediate producers consider the price of labor, capital and pollution discharge when choosing a production technique. The price of labor, w^e , measures the wage per effective labor unit, thereby accounting for labor quality differences across countries. The rental price of capital is given by r . If firms were

¹⁹ It is common to assume that pollution intensity rises with the capital intensity of production. Copeland and Taylor (2003) provide some evidence for the case of SO₂.

²⁰ As in Copeland and Taylor (1994), we restrict $D \leq IE$ for $I > 0$ to ensure that output is bounded above for a given labor input. See Copeland and Taylor for further discussion.

unregulated, they would always choose to discharge as much as possible to economize on labor. However, China levies a pollution tax, \mathbf{t} , according to (2), and this tax is effective in the sense that firms abate some pollution. Given these factor prices, the firm's labor and discharge combination that satisfies cost minimization is:

$$\frac{w^e}{\mathbf{t}} = \left(\frac{1 - \mathbf{a}(z)}{\mathbf{a}(z)} \right) \frac{D}{E}. \quad (7)$$

Because (7) implies that the parameter $\mathbf{a}(z)$ determines how pollution discharge varies among intermediates producers, $\mathbf{a}(z)$ provides a measure of pollution intensity. We can order the intermediates in order of decreasing pollution intensity to obtain $\mathbf{a}'(z) < 0$.

To determine the pattern of trade between China and the advanced countries, we examine how unit production costs vary across intermediates. The unit cost of producing one unit of input x in country i is given by

$$c(w_i^e, \mathbf{t}_i, r_i; z) = \mathbf{k}(z) w_i^{e(1-\mathbf{a}(z))q} \mathbf{t}_i^{\mathbf{a}(z)q} r_i^{1-q}, \quad (8)$$

where $\mathbf{k}(z)$ is an industry-specific constant. Input z is produced in an Advanced country if

$$c(w_A, \mathbf{t}_A, r_A; z) < c(w_C, \mathbf{t}_C, r_C; z).$$

We assume that labor in the Advanced bloc has high human capital levels and, thus, it is more productive than labor in China. The pollution tax levied in the Advanced countries exceeds the rate set in China, such that $\frac{w_A}{\mathbf{t}_A} < \frac{w_C}{\mathbf{t}_C}$. Given these relative factor prices, and assuming for the moment that rental rates

are the same in both countries, input z would be produced in the Advanced bloc if

$$\mathbf{w} \equiv \frac{w_A^e}{w_C^e} \leq \left(\frac{\mathbf{t}_C}{\mathbf{t}_A} \right)^{\mathbf{a}(z)(1-\mathbf{a}(z))} \equiv T(z). \quad (9)$$

With $\mathbf{t}_A > \mathbf{t}_C$, due to the higher real per capita income in the Advanced bloc and $\mathbf{a}'(z) < 0$, T must be

increasing in z . The Advanced bloc's cost advantage decreases as the pollution intensity of production increases.

Now we assume that the rental rate of capital is not the same in both countries and that instead, $r_A < r_C$. Because capital's cost share is the same across all goods, this rental differential lowers the cost of production in the Advanced countries across the full range of intermediates. To consider an equilibrium with some trade in intermediates, we assume that despite its lower rental rate the Advanced bloc has a cost disadvantage for some intermediate input z^* , defined as that input for which

$$c(w_A^e, \mathbf{t}_A, r_A; z) = c(w_C^e, \mathbf{t}_C, r_C; z).$$

Figure 5 shows the minimum cost locus for China as CC and for the Advanced bloc as AA.²¹ The absolute slopes of the loci are indeterminate, but they are upward sloping. Given our assumptions about comparative factor prices, China has lower costs than the Advanced bloc in the range of inputs indexed by $z \in [0, z^*)$ while the Advanced bloc has lower costs in the range $z \in (z^*, 1]$.

The pollution intensity of this fragmented sector depends on which inputs China produces; that is, it depends on the value of z^* . Based on the production functions (6), total discharge from the fragmented sector is

$$D = \int_0^{z^*} D(z) dz = \int_0^{z^*} \frac{\mathbf{a}(z)p(z)x(z)}{\mathbf{t}} dz \quad (10)$$

For simplicity, we assume that demand by the final good producer for each input is a constant share of total world expenditure and that, as a small country, China has a negligible impact on world income.²² Using this assumption, $p(z)x(z) = \mathbf{j}(z)I^W$, in equation (10) leads to an expression for the pollution intensity of the fragmented sector:

²¹ Feenstra and Hanson (1996) introduce a similar diagram to illustrate the fragmentation of production between the United States and Mexico.

²² Copeland and Taylor (1994) also assume that budget shares are constant in their model, but they consider two countries large enough to affect international markets.

$$e_x = \int_0^{z^*} \frac{D(z)}{p(z)x(z)} \frac{p(z)x(z)}{\int_0^{z^*} p(z)x(z)dz} dz = \frac{1}{\int_0^{z^*} p(z)x(z)dz} \int_0^{z^*} \frac{\mathbf{a}(z)\mathbf{j}(z)I^W}{t} dz \quad (11)$$

Equation (11) allows us to express the pollution intensity of the processing sector as a function of the pollution tax, t , and the critical value, z^* . Note that an increase in the critical value, z^* , reduces the average pollution intensity of the export processing sector because $\mathbf{a}(z)$ is a decreasing function of z . We also note that an increase in z^* reduces the pollution intensity of the inputs imported in this sector for processing. Thus, when the range of inputs produced in China expands, the pollution intensity of both exports and imports in this sector declines.

As discussed above, the critical value, z^* depends on the cost of intermediates production in China, $c(w_C^e, \mathbf{t}_C, r_C; z)$. Therefore, z^* depends on all determinants of factor prices for the processing sector. These determinants are the terms of trade and the level of trade frictions, the determinants of the pollution tax rate, and all factor endowment. As discussed in a previous sector, foreign investment has been skewed toward those sectors that process and assemble imported intermediates. Therefore, we separate the capital stock into domestic (K^d) and foreign owned capital (K^f), allowing us to express the pollution intensity of the export processing sector as:

$$e_x = e_x(K^d, K^f, H, L, p^*, \mathbf{d}, R). \quad (12)$$

The pollution intensity of the whole export bundle is a weighted average of the pollution intensity of ordinary exports and the pollution intensity of processing exports. Using (5) to express the pollution intensity of ordinary exports and (12) to express the pollution intensity of processing exports and letting S_x denote the share of total exports that are processing exports, the pollution intensity of China's trade bundle is:

$$e = S_y e_y + S_x e_x = e_y + S_x (e_x - e_y) = e(K^d, K^f, H, L, p^*, \mathbf{d}, R, S_x), \quad (13)$$

where we have used the fact that $S_y + S_x = 1$. Because we have assumed that $e_x < e_y$, an increase in the processing share of exports obviously reduces overall export pollution intensity, *ceteris paribus*.

Foreign capital flows primarily to the export processing sector, reducing its cost of capital. Figure 5 can be used to illustrate the effect of this capital inflow on China's input competitiveness. At constant wages and pollution tax, the curve labeled CC shifts down, causing z^* to rise from z_1^* to z_2^* . With the pollution tax unchanged, there is no change in the pollution intensity of any intermediate. However, the capital inflow pulls labor into the processing sector, raising its share in exports. Moreover, because China now produces intermediates that are less pollution intensive than any it produced before, the average pollution intensity of China's export processing exports falls.²³ Likewise, the pollution intensity of China's imports falls because China now imports a narrower set of inputs and this set is, on average, cleaner than before.

Foreign investment may reduce export pollution intensity through another channel, which we have not formally modeled, even if we hold the processing share of exports fixed. Foreign investment often involves the use of new capital equipment and new production techniques. In particular, investment from high-regulatory-standard countries may transfer new pollution control methods to the host country as investors use technology and techniques that they have developed within the context of stringent pollution regulation.²⁴ If foreign investors bring this sort of "technique effect" with them, the pollution intensity of China's exports should be negatively associated with the level of foreign capital, even when the share of processing exports is held constant.

V. Estimating the Determinants of the Pollution Intensity of China's Manufacturing Trade

To test the determinants of the pollution intensity of China's trade, we begin with the simple model in which there is no fragmentation and FDI plays no distinct role. We specify the reduced form model in equation (5) as:

²³ There will also be feedback effects, which we do not discuss here. First, increased foreign investment may raise domestic wages, but this wage effect cannot overturn the direct effect of foreign investment. Second, higher real per capita income implies a higher pollution tax, reinforcing the direct effect by further reducing pollution intensity.

²⁴ This possibility is consistent with evidence presented in Dean, Lovely, and Wang (2006) on the location decisions of foreign investors. While provincial variation in pollution taxes influenced the location of Chinese investors, no effect was found for OECD investors.

$$\ln e_{it}^j = \mathbf{a}_{it} + \mathbf{b}_1 \ln K_{it} + \mathbf{b}_2 \ln E_{it} + \mathbf{b}_3 \ln p_{it} + \mathbf{b}_4 \ln \mathbf{d}_{it} + \mathbf{b}_5 \ln R_{it} + \mathbf{b}_6 trend_t + \mathbf{e}_{it}$$

where j is exports or imports, i is pollutant, and t is time. Since the log pollution intensity of exports (imports) in figure 3 follows a linear trend, one could estimate (5) using our panel data on four pollutants over the period 1995-2004, with pollutant-specific fixed effects,. Alternatively, we can wash out the impact of the trend and the need for fixed effects by expressing (5) as:

$$\Delta \ln D_{it}^j = \mathbf{g}_{it} + \mathbf{b}_1 \Delta \ln K_{it} + \mathbf{b}_2 \Delta \ln E_{it} + \mathbf{b}_3 \Delta \ln p_{it} + \mathbf{b}_4 \Delta \ln \mathbf{d}_{it} + \mathbf{b}_5 \Delta \ln R_{it} + \mathbf{h}_{it} \quad (5)'$$

where Δ indicates first difference.

Equation (5)' is estimated using pooled data on COD, SO₂, smoke and dust intensity of exports (imports) at the national level, from 1995-2004. After differencing, this yields a small panel of 36 observations. The estimation method is GLS with cross-section weights, to correct for pollutant-specific heteroskedasticity. Two caveats should be noted regarding this small panel data sample. First, there is some evidence of contemporaneous correlation across the pollutants in the sample, particularly with respect to the three air pollutants. Unfortunately, the sample is too small to also correct for contemporaneous correlation in all specifications. Second, it might be reasonable to split the sample, since the pollution intensity of trade with respect to water pollution may respond differently than with respect to air pollution. Again, the limited sample size prevents us from exploring this possibility.

Data on all explanatory variables except the tariff and trade variables are from the World Bank, *World Development Indicators, 2007*. These data are shown in Table 3. The log difference in the capital stock is proxied by gross capital formation (% of GDP), while the log difference in the total labor force and in real GDP per capita are calculated directly from the data.²⁵ For the moment, we do not differentiate investment by source, nor labor supply by skill level. To proxy the log difference in relative prices, we use the difference in China's net barter terms of trade, where the latter is defined as the ratio of the export price index to the import price index, measured relative to the base year 2000. The data used to calculate the log difference in tariffs are China's simple average MFN tariffs (ad valorem equivalent) taken from the UNCTAD TRAINS

²⁵ GDP per capita is in constant 2000 US dollars.

database, via WITS.²⁶

Table 4 presents the results of estimation of equation (5)' for exports in column (1). These results support some of the predictions discussed above. Assuming all China's exports are "ordinary," an increase in the capital-labor ratio increases the pollution intensity of exports, suggesting that capital and pollution may be complements in production. Increased stringency in environmental regulations (proxied by growth in real GDP per capita) does reduce the pollution intensity of exports, though the impact is not significant. Trade liberalization appears to be favorable for China's environment. A 1% drop in China's average tariff is associated with a 1.6% drop in the pollution intensity of exports. Since China's tariffs actually fell by about 75% during this period, this suggests that trade reform contributed significantly to China's cleaner trade. In addition, China's entrance into the WTO in 2001 also seems to have been associated with a significant reduction in the pollution intensity of China's exports. Finally, though the impact of a change in the terms of trade is indeterminate in theory, here an improvement in the terms of trade is associated with increased pollution intensity of exports. The parallel results for the pollution intensity of imports are shown in Table 5, column (1). While the results for trade barriers and entrance into the WTO are similar to that of exports, the results for other variables are much weaker.

Composition effects and fragmentation

Moving beyond the simple model, we incorporate both ordinary and fragmented exports, as in the reduced form model in equation (13). Equation (13) suggests that changes in overall pollution intensity will be explained not only by the changing pollution intensity of ordinary exports, as in (5)', but by growth in the share of fragmented exports and changes in that subsector's pollution intensity. Our proxy for the share of exports (imports) which are fragmented is the share of processing exports (imports) in total exports (imports). This variable is calculated directly from the trade data from China Customs; it includes both exports (imports) designated as processing and assembly and those designated as processing with imported materials. We begin

²⁶ TRAINS has no Chinese tariff data for 1994-1995 or 2002. The simple average MFN tariff data for 1994-5 (with no AVE correction) was taken from Zhang et al.(1998), and for 2002 (with no AVE correction) was taken from the WTO (2006).

by treating the processing share as exogenous, and simply add the change in this share to equation (5)'.

The results of this estimation (column (2) of table 4) support the idea that increased fragmentation has reduced the pollution intensity of China's exports. An increase in the share of processing exports by a percentage point reduces the pollution intensity of China's exports by about 0.02%. The share of processing exports actually grew by about 6% during this time period, implying a larger impact than the small elasticity might suggest. The inclusion of the export processing share also strengthens the magnitude and significance of factor endowments and environmental stringency in explaining the growth in pollution intensity over time. The parallel results for imports (table 5, column (2)) are even more striking. The impact of an increase in the share of processing imports on the pollution intensity of China's imports is much larger and more significant (compare table 4 and 5, column (2)). The inclusion of import processing share also dramatically strengthens the significance of all other explanatory variables compared to the case where the fragmented sector was ignored (table 5, column (1)).

However, the size of the fragmented sector is most likely endogenous. Clearly changes in trade frictions and factor endowments influence the size of the processing export share. Trade barriers on imports in highly fragmented sectors have fallen over this time period.²⁷ China's entrance into the WTO has also meant more favorable access for China's ordinary and fragmented exports in other WTO members' markets. As discussed above, growth in foreign investment is predicted to raise the processing share of exports. Similarly, if export processing is more human-capital intensive than ordinary export processing, growth in the relative supply of human capital will raise the share of resources devoted to export processing. To account for this endogeneity, we re-estimate equation (13)', using instrumental variables.²⁸ The instrumented results (column (3)) now show much stronger evidence that growth in the share of fragmented exports leads to cleaner exports. The elasticity of pollution intensity with respect to processing export share has more than

²⁷ For example, the WTO (2006) reports that average tariffs on electronic and communications equipment imports fell with accession to the WTO. In April, 2003 China joined the WTO Information Technology Agreement, and 258 tariff lines at the HS 8-digit level became subject to zero tariffs. Import licenses and quotas on certain products were also been removed.

²⁸ The instruments for processing export (import) share include all the other variables in the equation and the share (lagged share) of processing imports in total imports.

tripled, and it is now highly significant. Similar findings (though less dramatic) appear in the parallel results for imports (table 5, column (3)).

Composition effects, technique effects, and FDI

Thus far we have not distinguished investment by source nor labor by skill. Yet, FDI plays a crucial role in fragmented trade. As argued above, an increase in FDI flows should reduce pollution intensity by increasing the share of processing exports and by increasing the critical value, z^* . Domestic capital, in contrast, flows primarily to the import-competing and ordinary export sectors. Thus, an increase in domestically-sourced investment pulls factors out of the export-processing sector, reducing the critical value z^* , and increasing the average pollution intensity of the export-processing sector.²⁹ Production shifts to the more highly polluting ordinary export sector. Therefore, we expect that an increase in domestic investment raises the pollution intensity of China's exports.

An increase in the relative supply of human capital acts, in the model, like a decrease in the Chinese effective wage. A decrease in w^e shifts the CC line down in figure 5, allowing China to compete successfully in production of more human-capital-intensive intermediate inputs. Thus, an increase in Chinese human capital is predicted to reduce the pollution intensity of China's exports. An increase in unskilled labor, on the other hand, is predicted to have the opposite effect.

The last two columns of table 4 show evidence that is certainly suggestive of the important role that increased FDI and increased human capital play in making Chinese exports cleaner. In column (4) of table 4, we present results for the instrumented estimation of (13)' again, but with investment split between domestically-sourced investment and FDI. FDI (% of GDP) is taken from the *World Development Indicators*.³⁰ Domestically-sourced investment (as a share of GDP) is calculated as the difference between gross capital formation and FDI. It is immediately evident that these two types of investment have opposite effects. As expected, increased FDI flows strongly reduce the pollution intensity of Chinese trade, while

²⁹ The CC line in Figure 5 shifts up when labor is pulled out of the sector and wages rise.

³⁰ These data closely parallel official Chinese data on utilized (or realized) FDI flows (% GDP) (see Annual FDI Statistics, www.fdi.gov.cn).

increased domestically-sourced investment does the opposite. Because the effects of FDI flows on the size of the fragmented sector are captured via the IV estimation, the coefficient on the FDI variable actually suggests evidence of cleaner exports due to a change in composition *within* the fragmented sector (an increase in z^*). It may also suggest that foreign investors bring greener technologies than their local counterparts, implying an additional favorable technique effect. Parallel results for imports (table 5, column (4)) are much weaker and show no such role for FDI.

Because of the small sample size, we are unable to test for distinct roles of investment by source and labor by skill simultaneously. However, some evidence suggestive of the importance of both is shown in column (5) of table 4. In this final regression, we include the ratio of FDI to domestically-sourced investment as well as growth in the ratio of skilled to unskilled labor. The latter is proxied by the share of the population with at least senior secondary education, relative to the illiterate share.³¹ The results in column (5) suggest that the pollution intensity of exports is strongly reduced by the relative growth of foreign investment and of skilled labor. This evidence is consistent with the notion that increased FDI flows expands the composition of fragmented exports to include cleaner intermediates and that more skill-intensive intermediates are cleaner. While the theory would suggest both these attributes should be true of imports as well, only the FDI results are borne out in table 5 (column (5)).³²

VI. Global Engagement and the Environment

By all accounts, China's rapid economic growth over the past 20 years has been accompanied by severe environmental degradation. While much of this deterioration can be attributed to growth in domestic consumption, the extent to which China's environment has been sacrificed so that it can serve as "the world's factory" is an important economic and moral question. To begin to address this issue, this paper provides new

³¹ Data on shares of population aged >6 years by educational attainment are from various issues of the China Statistical Yearbook. Data for the year 1995 are from Cao (2000), page 4.

³² The results for the impact of the ratio of skilled to unskilled labor on the pollution intensity of imports appear to be highly sensitive to the lag chosen. More data are required to determine how illustrative they really are.

evidence on trends in industrial pollution intensity, changes in the pollution intensity of Chinese trade, and the influence of foreign investment and production fragmentation on the pollution content of Chinese exports and imports. Contrary to the expectations of many commentators, we find that deeper global engagement has reduced the implicit environmental cost of Chinese production and trade.

Using official Chinese environmental data on air and water pollution from SEPA, we find that industrial emissions of primary pollutants have slowed or fallen over the last decade while trade has grown. Relative to 1995 levels, manufacturing trade increased almost 300% by 2005, while annual industrial emissions of COD, smoke, and dust declined by 56%, 46%, and 40%. Only industrial emissions of SO² rose after 1999, as they were 17.5% higher by 2005. As noted by Naughton (2007, p. 495), the abatement of waste from large factories has been a relatively positive part of China's environmental record and the stabilization of waste while output has grown sharply represents a significant achievement in its development.

Using emissions data compiled from Chinese Environmental Yearbooks, we present new evidence on the pollution intensity of Chinese industrial production. Tracking changes in these pollution intensities over time reveals surprising trends. Across all four pollutants, we find that the pollution intensity of almost all sectors has fallen since 1995. This finding suggests that China has benefited from a positive "technique effect," as emissions per real dollar of output have fallen across a wide range of industries. Suggestively, a review of trends in Chinese trade patterns reveals that China's trade appears to be shifting toward relatively cleaner sectors over time. In particular, the share of exports accounted for by textiles and leather products has fallen while the share accounted for by office and computing, and communications equipment has grown dramatically. These growth sectors are characterized by very low air and water pollution intensities, and by high shares of processing trade, indicating the substantial presence of two-way trade in production "fragments."

Linking the industrial pollution intensities to detailed trade statistics from China Customs, yields a weighted average pollution intensity for China's manufacturing exports (imports) for each year in the period 1995 to 2005. Contrary to popular expectations, which emphasize the migration of dirty industries to poor nations, we find that Chinese exports are less water pollution intensive, and generally less air pollution

intensive, than Chinese imports would be if produced domestically. Moreover, both Chinese exports and imports are becoming cleaner over time. Holding the pollution intensity of production constant in a counterfactual experiment, we find that changes in the composition of trade over the decade account for some of the trend toward cleaner trade, although a substantial share of the decline remains attributed to changes in production techniques. Finally, we find that processing trade is cleaner than ordinary trade.

The weight of this evidence certainly suggests that the increased concentration of Chinese trade in highly fragmented industries has led to composition and technique effects which are favorable toward China's environment. Drawing on Copeland and Taylor (1994, 2003), we present a simple model of production and trade that leads to a reduced form equation for the pollution intensity of Chinese trade. Explicitly incorporating a role for fragmented trade (drawing on the work of Feenstra and Hanson (1996)), yields a set of key determinants of the pollution intensity of trade: Chinese domestic factor endowments, foreign investment, the terms of trade, trade frictions, per-capita real income, and the share of trade in fragmented sectors, where this share is also influenced by the other key determinants. In theory, increased FDI inflows not only increases the size of the fragmented sector, but also reduces its average pollution intensity.

Econometric evidence from instrumental variables estimation strongly supports the role of processing trade in explaining the drop in the pollution intensity of Chinese exports and imports over time. This suggests that there is indeed a favorable composition effect generated by the increased importance of fragmentation in Chinese trade. The evidence also suggests that, controlling for the size of processing exports, FDI inflows contribute to cleaner exports. This supports the idea that increased FDI may change the composition of the fragmented sector itself toward relatively cleaner intermediate goods, and may also bring greener technology to the fragmented sector.

In the Five-Year Plan for 2006-2010, the Chinese authorities call for a reorientation of their economic growth model toward environmental sustainability. How China will achieve the dual goal of economic growth and reduced environmental degradation is far from clear. Trade and foreign investment has fueled much of China's trade boon and so it is natural to ask whether China's unique brand of global engagement needs to be radically altered to move its development path in the desired direction. The new data analyzed in

this paper suggests that, at least provisionally, the answer to this question is “no.” Industrial pollution intensity has already stabilized and, in many industries, has already begun to decline. Looking specifically at the bundle of goods China trades with the world, we find that, contrary to what might have been expected, foreign investment and integration into global production networks has reduced the environmental cost of China’s growth.

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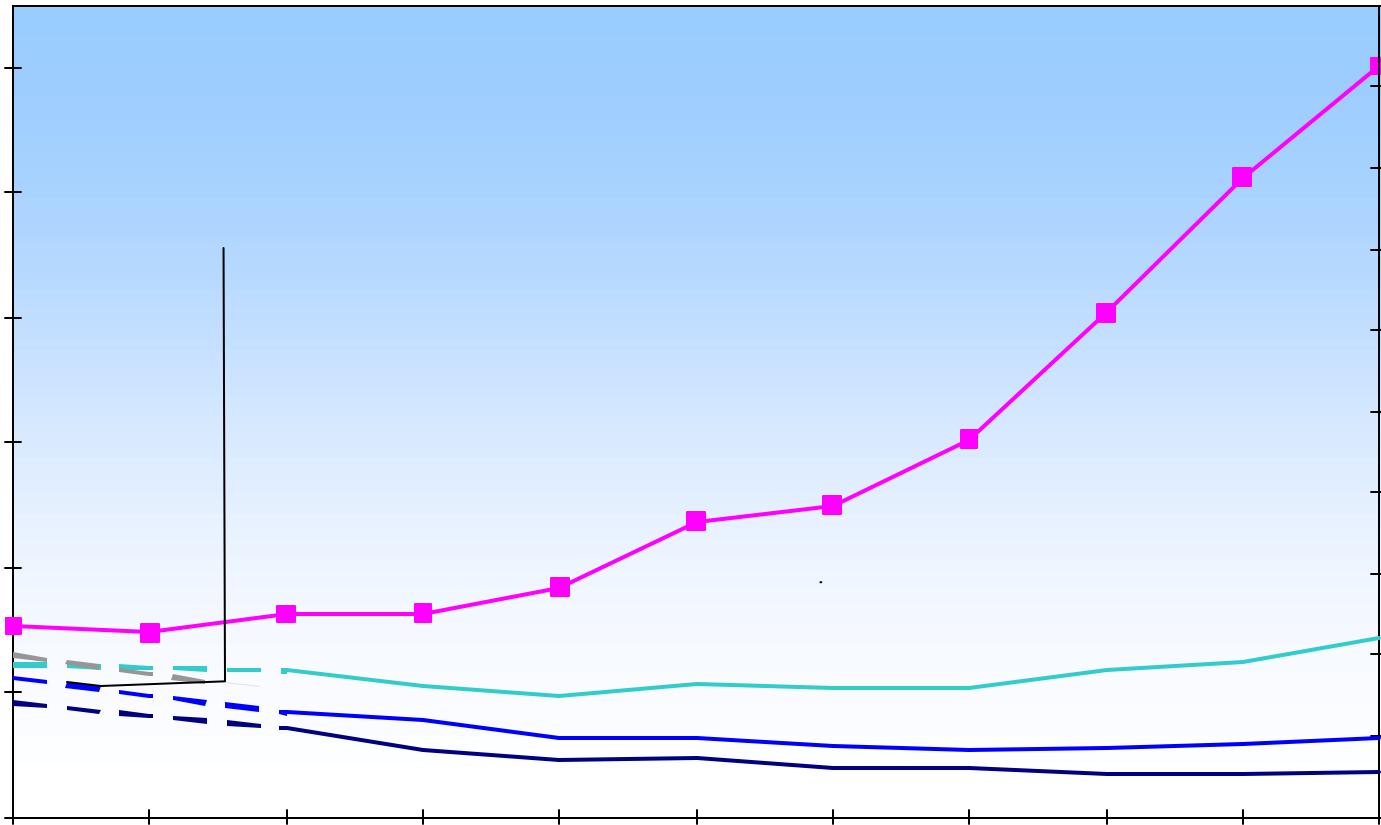


Figure 2. China's Trade and Industrial Emissions
(Index, 1995=100)

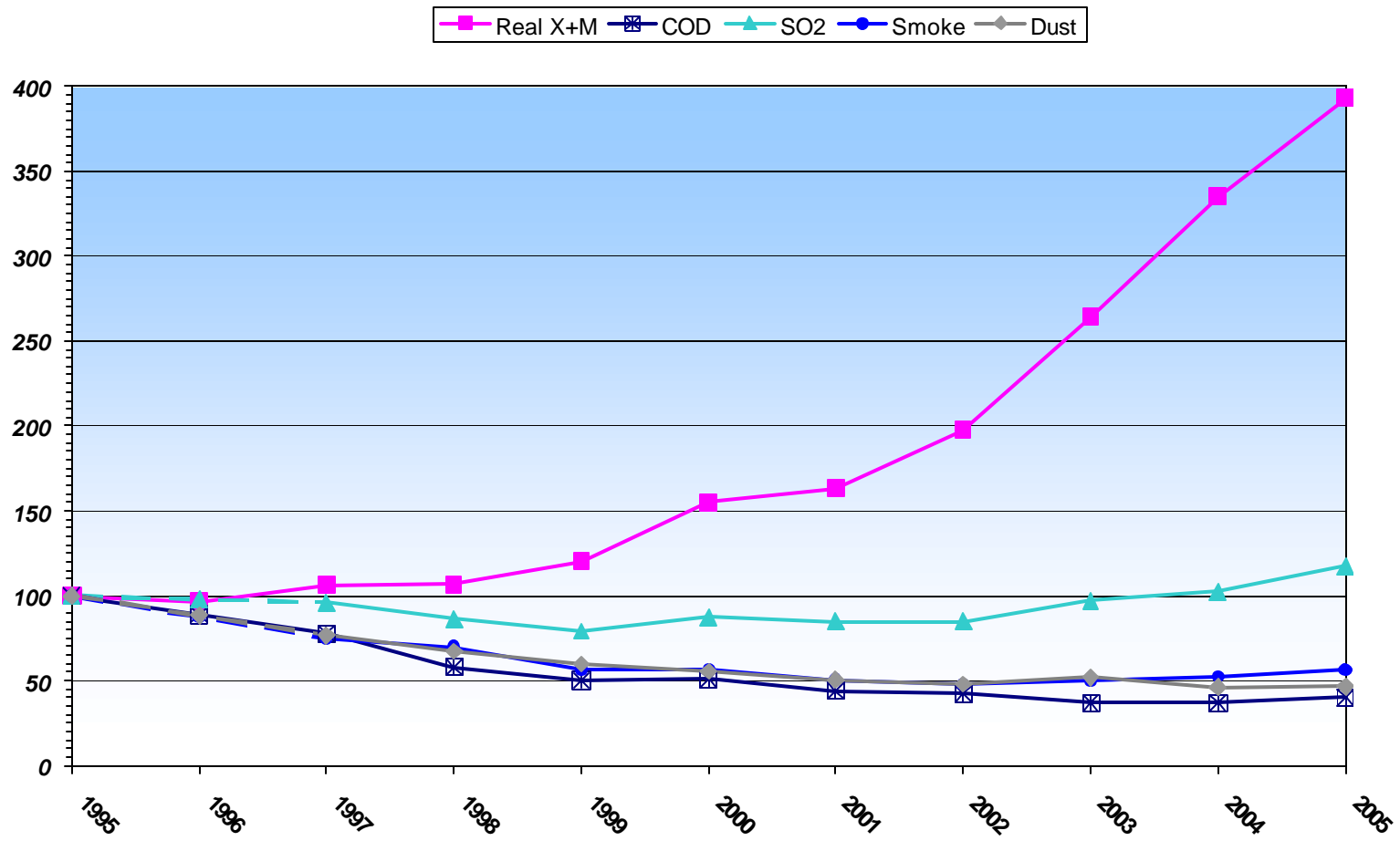


Figure 3.

The Pollution Intensity of China's Trade: 1995-2004

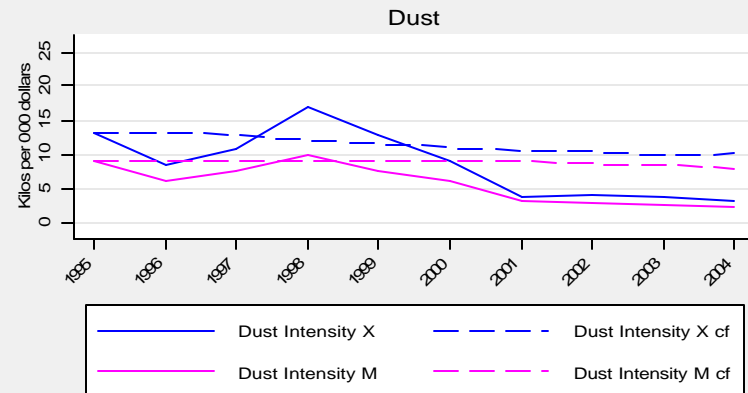
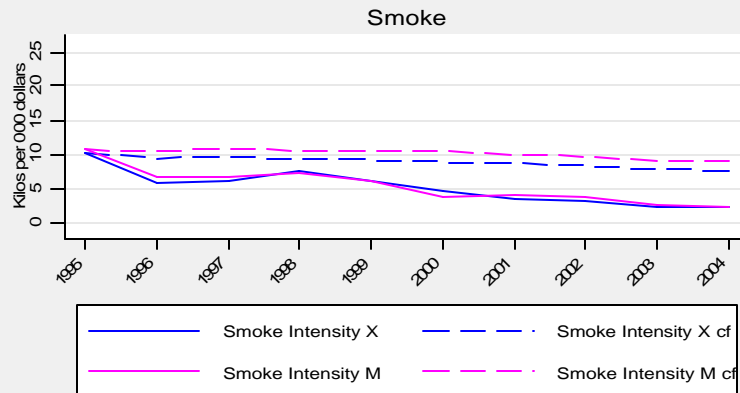
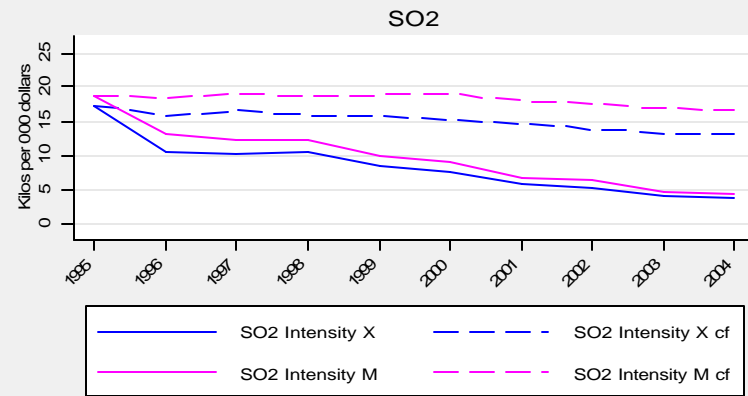
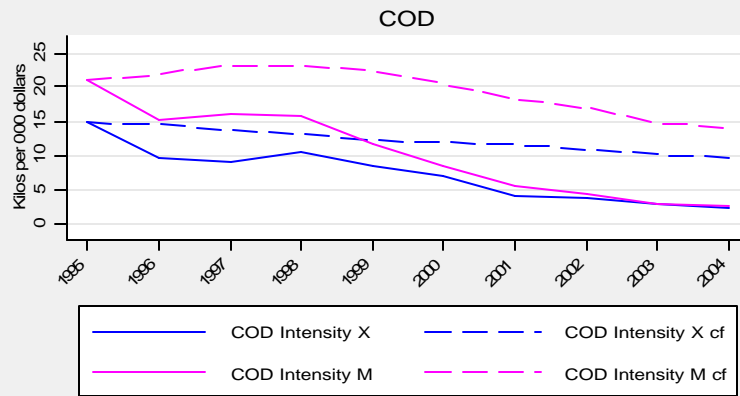


Figure 4.

The Pollution Intensity of China's Processing Trade: 1995-2004

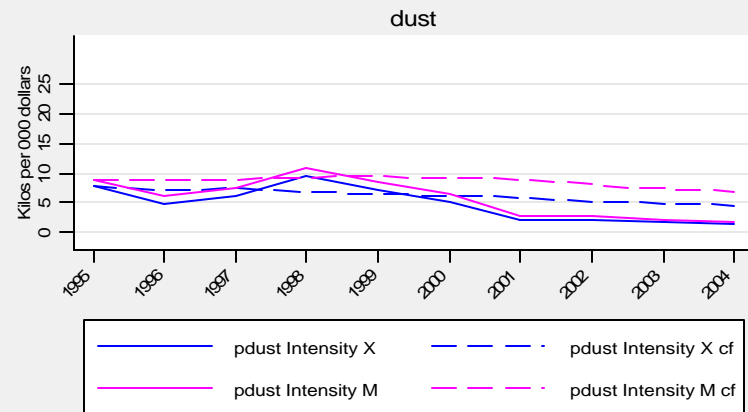
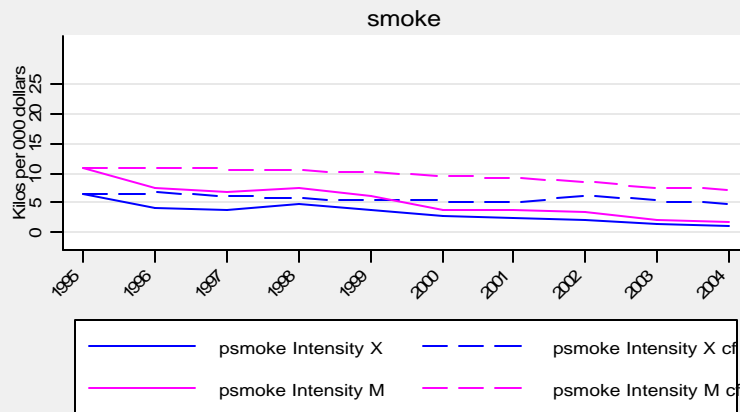
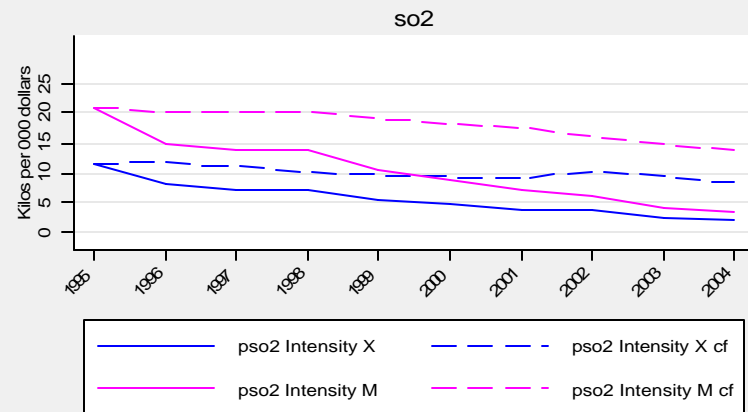
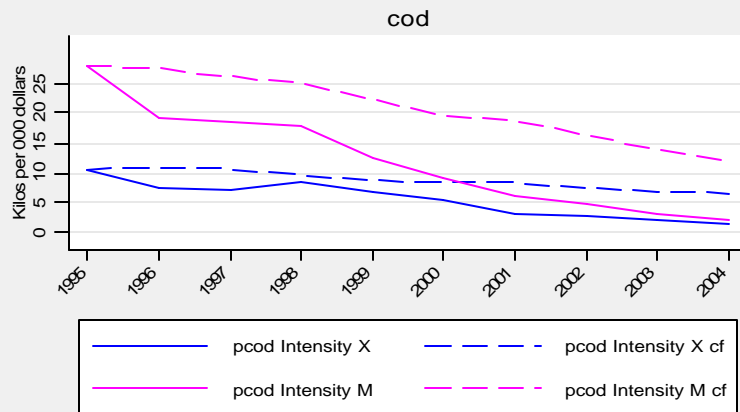


Figure 5: FDI Expands Range of Export Processing Activities Performed in China

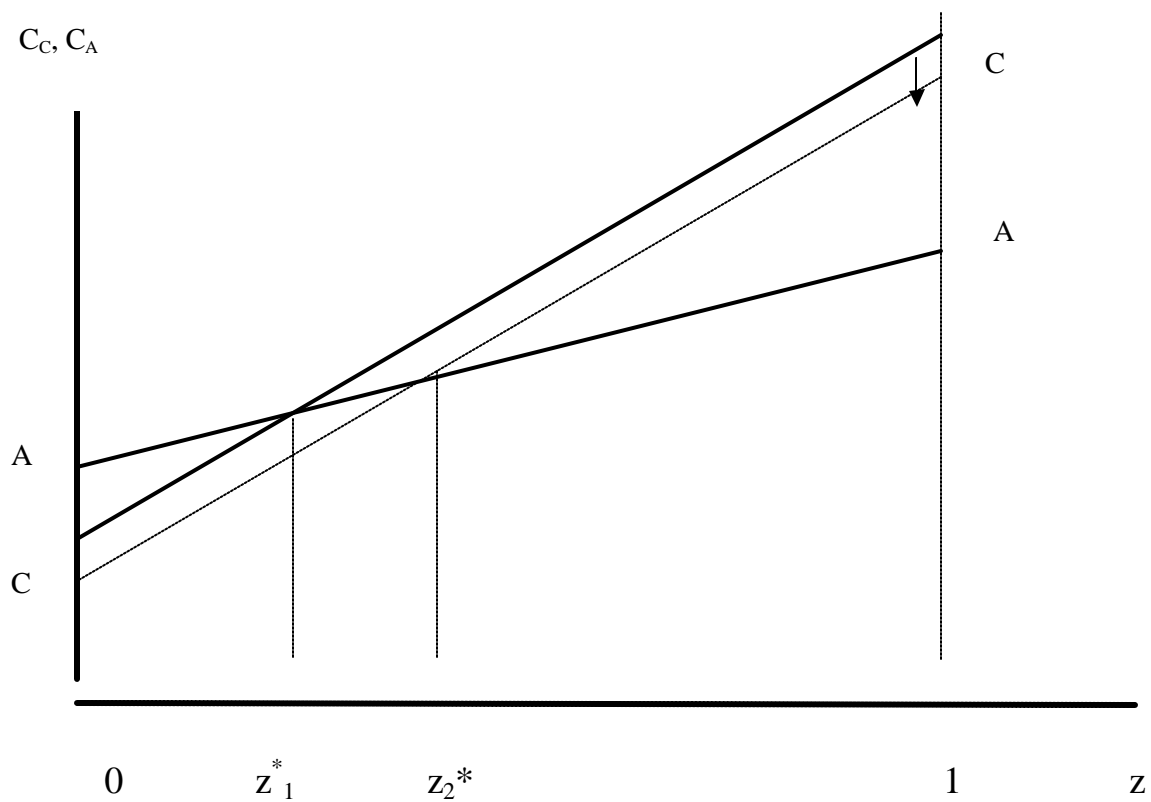


Table 1. The Composition of China's Trade, 1995 and 2004

ISIC Rev. 3 Two Digit Sector	China Mfg Exports				China Mfg Imports			
	Share of Total Mfg Exports (%)		Processing Exports as a Share of Mfg Exports (%)		Share of Total Mfg Imports (%)		Processing Imports as a Share of Mfg Imports (%)	
	1995	2004	1995	2004	1995	2004	1995	2004
15 Food Products and Beverages	5.5	2.6	24.4	31.0	4.9	2.4	45.2	26.2
16 Tobacco	0.7	0.0	26.6	2.9	0.3	0.0	0.3	0.3
17 Textiles	13.8	8.2	32.3	25.7	9.1	3.1	97.2	90.7
18 Wearing Apparel	14.2	8.6	54.4	31.0	0.8	0.3	96.5	73.0
19 Leather Shoes	7.3	4.1	72.7	47.0	1.9	0.8	98.6	85.9
20 Wood	1.5	1.0	14.8	19.4	1.0	0.5	44.4	59.9
21 Paper	0.6	0.4	42.3	59.8	2.5	1.7	66.1	38.2
22 Printing	0.1	0.1	79.5	54.3	0.0	0.0	55.0	35.2
23 Coke and Petroleum	1.3	1.5	26.2	24.4	2.2	2.6	9.1	2.7
24 Chemicals	6.8	4.8	21.0	25.3	15.8	14.3	53.3	33.3
25 Rubber and Plastics	2.7	2.7	71.7	62.7	1.8	1.7	83.0	56.1
26 Non-metallic minerals	2.3	1.7	14.7	17.2	0.8	0.7	40.1	48.5
27 Basic metals	5.2	4.1	56.1	27.5	8.8	8.4	52.3	40.1
28 Fabricated metals	3.4	3.5	36.9	25.5	1.8	1.3	43.2	37.7
29 Machinery	4.7	7.2	45.7	48.2	20.7	12.8	3.8	7.8
30 Office and Computing Machinery	3.5	15.1	94.7	95.8	2.4	6.2	66.8	50.3
31 Electrical Machinery	5.1	5.8	69.9	62.4	5.1	6.0	50.7	52.6
32 Communications Equipment	7.8	15.7	85.6	86.0	10.4	23.0	59.8	71.9
33 Medical, Precision and Optical Instruments	2.9	3.0	80.5	76.2	3.6	8.6	42.8	57.0
34 Motor vehicles	1.4	2.1	73.6	59.8	2.5	3.3	4.2	2.1
35 Transport equipment	1.5	1.7	59.6	53.3	2.5	1.4	7.8	4.5
36 Furniture and Other Mfg.	7.9	6.3	68.6	59.7	1.1	0.6	72.2	57.4

s e g a r o 15 c a Food Products and Beverages t 4 9 8 0 T 5 7 . 0 1 0 6 5 2 . 0 - c T 7 8 7 7 . 0 T 4 W r o o F 3 0 . 0 3 1 . 0 0
 D (c T 9.94 9 7 2.60 . 0 1.81 c T 0.13 7 8 1.41 5 7 .0.60 T 4.57 W r o 0.63 F

1995			
COD	SO2	Smoke	Dust

2004			
COD	SO2	Smoke	Dust

	(1) Equ. (5)'		(2) Equ. (13)		(3) Equ. (13) IV		(4) Equ. (13) IV		(5) Equ. (13)IV	
<i>Variables in log difference unless otherwise noted.</i>	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴
<i>Gross Capital Formation²</i>	0.04*	2.35	0.06**	2.89	0.13**	2.98				
<i>Domestic Investment²</i>							0.13**	4.89		
<i>FDI²</i>							-0.48**	-2.96		
<i>Ratio of FDI to Dom. Inv.²</i>									-0.11**	-3.49
<i>Ratio of Skilled to Unskilled Labor³</i>									-0.02**	-4.46
<i>Labor Force</i>	-0.38*	-2.02	-0.43*	-2.21	-0.63†	-1.99	0.17	0.70		
<i>Real GDP p.c.</i>	-0.05	-1.08	-0.09†	-1.78	-0.23*	-2.31	-0.38**	-4.02	-0.10*	-1.84
<i>Terms of Trade³</i>	0.05**	6.01	0.05**	5.74	0.05**	3.60	0.13**	5.34	0.10**	6.75
<i>Average Tariff</i>	1.51**	5.89	1.28**	4.18	0.54	0.87	0.81**	2.72	1.38**	5.43
<i>WTO Dummy</i>	-0.53**	-8.26	-0.51**	-7.64	-0.47**	-4.37	-0.92**	-6.66	-1.00**	-9.16
<i>Processing Exports Share³</i>			-0.02†	-1.71	-0.09*	-2.50	-0.03**	-3.01	-0.03**	-3.09
<i>Constant</i>	-0.71	-1.68	-1.06*	-2.24	-2.26*	-2.57	0.67	1.00	2.44**	3.12
Obs.	36		36		36		36		36	
Weighted Adj. R² ⁴	0.77		0.76		0.50		0.82		0.83	
Weighted F-statistic ⁴	20.04		16.72**		17.23**		21.08**		24.28**	
Wald test⁵ (?²)	120.26**		117.03**							

Notes: **, * and † indicate statistical significance at the 1%, 5% and 10% levels, respectively.

¹ Dependent variable is log difference of pollution intensity of exports. All regressions are GLS with panel-specific weights to correct for pollutant-specific heteroskedasticity.

² Expressed as share of GDP.

³ Expressed as difference between value in period t and period t-1.

⁴ Eviews output gives weighted adjusted R² and F-statistics, where the weights adjust for the cross-section weights. Eviews also gives t-statistics rather than z-statistics.

⁵ Test s null hypothesis that all coefficients (except the constant) are simultaneously equal to zero.

	Equ. (5)'		Equ. (13)		Equ. (13) IV		Equ. (13) IV		Equ. (13)IV	
<i>Variables in log difference unless otherwise noted.</i>	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴	Coeff.	t-stat ⁴
<i>Gross Capital Formation</i> ²	0.03	1.05	0.13**	3.69	0.18**	3.05				
<i>Domestic Investment</i> ²							0.15**	3.79		
<i>FDI</i> ²							0.53	1.31		
<i>Ratio of FDI to Dom. Inv.</i> ²									-0.22**	-2.65
<i>Ratio of Skilled to Unskilled Labor</i> ³									0.13**	3.17
<i>Labor Force</i>	-0.12	-0.49	-0.38†	-1.77	-0.50†	-1.98	-0.90	-1.57		
<i>Real GDP p.c.</i>	-0.01	-0.22	-0.19**	-2.85	-0.26**	-2.61	-0.10	-0.86	-0.96**	-3.03
<i>Terms of Trade</i> ³	0.05**	4.24	0.06**	6.19	0.07**	5.50	0.02	0.48	0.05**	4.07
<i>Average Tariff</i>	1.20**	3.50	1.19**	4.34	1.20**	4.17	1.42**	3.93	-5.08**	-2.76
<i>WTO Dummy</i>	-0.35**	-4.05	-0.45**	-5.87	-0.49**	-5.17	-0.24**	-1.04	-2.62**	-3.57
<i>Processing Imports Share (lagged)</i> ³			-0.05**	-3.69	-0.07**	-2.81	-0.07**	-2.81	-0.03**	-2.21
<i>Constant</i>	-0.65	-1.15	-2.81**	-3.74	-3.65**	-3.11	-4.97**	-2.16	9.00**	2.91
Obs.										
Weighted Adj. R² ⁴	0.42		0.59		0.56		0.59		0.55	
Weighted F-statistic ⁴	5.29**		8.23**		6.16**		7.40**		6.44**	
Wald test⁵ (?²)	31.74**		57.60**							

Notes: **, * and † indicate statistical significance at the 1%, 5% and 10% levels, respectively.

¹ Dependent variable is log difference of pollution intensity of exports. All regressions are GLS with panel-specific weights to correct for pollutant-specific heteroskedasticity.

² Expressed as share of GDP.

³ Expressed as difference between value in period t and period t-1.

⁴ Eviews output gives weighted adjusted R² and F-statistics, where the weights adjust for the cross-section weights. Eviews also gives t-statistics rather than z-statistics.

⁵ Tests null hypothesis that all coefficients (except the constant) are simultaneously equal to zero.

Appendix

Construction of the pollution intensities of Chinese manufacturing industries, 1995-2004

Data on emissions of COD, SO₂, smoke and dust, as well as the current value of output of the sampled enterprises at the industry level, were compiled by the authors from the *Chinese Environmental Yearbooks* (Chinese editions) and the *China Statistical Yearbook on Environment* (dual language, 2000, 2005 and 2006). Emissions data are originally in tons and output in 1000 current yuan. They are available by the 2-digit “divisions” in the Chinese industrial classification system for the industrial sector, which includes Mining (6 divisions), Manufacturing (30 divisions) and Distribution of Electricity, Water and Gas (3 divisions). Pollution intensities were calculated as emissions (in kilos) per thousand real yuan (1995 yuan). Output was deflated using the manufacturing producer price index (*China Statistical Yearbook, various issues*). These pollution intensities are shown for Manufacturing and for the Distribution of Electricity, Water and Gas, by division (GB/T 4754-2002), in table A.1.

Change in Chinese Industrial Classifications

Prior to 2001 Chinese industrial data were classified using GB/T 4754-1994. From 2001 onwards industrial data are classified using GB/T 4754-2002. In both classifications, manufacturing has 30 2-digit “divisions.” Using the official Chinese concordance, we compared the two classifications and found only two changes in manufacturing divisions.¹ First, the 1994 division 39 (weapons and ammunition mfg.) became part of 2002 division 36 (special equipment mfg.).² We address this change under aggregation issues below. Second, the 2002 division 43 (“waste recycling”) was added. This division was not part of manufacturing in the previous period. Therefore, we dropped it from the analysis.

Aggregation and Missing Data

In the published emissions and output data from 1995-2000, several divisions are aggregated together. Divisions 13-16 are grouped as “Food, Beverages and Tobacco,” divisions 35-41 are grouped as

¹ The 4-digit “classes” within each 2-digit division remained essentially unchanged. There were fewer classes in total in the 2002 classification, largely due to merges of classes within the same division.

² The remaining 2002 division codes were renumbered accordingly. Thus, 1994 division 40 corresponds to 2002 division 39, 1994 division 41 corresponds to 2002 division 40, etc.

“Machine, Electric Machinery & Electronic Equipment Mfg.,” and divisions 44-46 are grouped as “Production and Supply of Electric Power, Gas, and Water. To disaggregate these grouped data, we first created corresponding groups for the years 2001-2004 by summing the appropriate division data. For each group, we calculated the average share of emissions of each pollutant attributable to each division within the group. We then applied these shares to the recorded group data in the earlier period. The group’s annual emissions data from 1995-2000 for each pollutant was multiplied by the corresponding average share to derive the missing annual emissions data for each division within that group. We followed a similar procedure to derive the missing output data for each division within each group.

For example, during 2001-2004, Food Production (14) was responsible on average, for about 16 % of annual COD emissions and about 17% of annual output of “Food, Beverages and Tobacco.” Therefore, for each year during 1995-2000, 16% of the recorded COD emissions and 17% of the recorded output for that group were allocated to Food Production.

This method assumes that the 2001-2004 relative trends in emissions of each pollutant and in output across divisions within a group apply during the earlier period. This is certainly plausible. However, it could mask any radical changes in technique or in composition within a group which took place in a single year.

Emissions and output data for 5 divisions during the 1995-2000 period are missing: Clothes, Shoes and Hat Manufacture (18), Timber Processing, etc. (20), Furniture Manufacturing (21), Cultural, Educational and Sports Articles (24), and Craftwork and Other (42). To fill in the missing data for the first three, we paired each missing division with a related division for which complete data were available: (18) with (17) textiles; (20) with (22) papermaking and paper products; (21) with (22). For each pair, we calculated the average ratio of emissions of each pollutant for the missing division relative to the complete division during 2001-2004. These ratios were then applied to the recorded data for the complete division in the earlier period. For each year of 1995-2000 we multiplied the complete division’s data by these average emissions ratios, to derive the annual emissions data for the missing division in that pair. We then followed a similar procedure to derive the output data for the missing

division.

For example, during 2001-2004 we found that the ratio of COD emissions for Clothes (18) relative to Textiles (17) averaged about 3.3%, while the ratio of SO₂ emissions averaged about 4.1%. Therefore, for each year during 1995-2000, we assigned values for division (18) COD and SO₂ emissions that were 3.3% and 4.1%, respectively, of the recorded data for division (17).

We were unable to find a related division to pair with (24) or (42). Therefore, these data are missing during 1995-2000.³ These missing data essentially impact our estimates of the pollution intensity of ISIC 36 (Furniture and other manufacturing, not elsewhere specified). Division (24) maps almost exclusively to ISIC 36. The classes in division (42) map to several 2-digit ISIC categories, but mostly to ISIC 36. These two divisions accounted for 76% (47% and 29%, respectively) of ISIC 36 exports in 1995, but declined in importance over the period. By 2000, they accounted for only 57% (45% and 12%, respectively), while furniture's share had roughly doubled (11% to 19%). Thus, while the pollution intensity of exports of ISIC 36 in our analysis during 1995-2000 is based nearly exclusively on the pollution intensity of furniture production, any bias this may introduce diminishes over these five years⁴

Emissions from township and village level enterprises (TVIEs)

Emissions data prior to 1998 were recorded only for industrial enterprises at the "county level and above." After the "*Investigation on Sources of Township Industrial Pollution*" (1997), it was found that township and village industrial enterprises (TVIEs) were accounting for a significant and growing percentage of emissions. Therefore, the emissions data included these enterprises from 1998 onwards. Because TVIE emissions data are unavailable at the sectoral level, the yearly industrial pollution intensities in 1995-1997 do not include TVIEs. Thus, the values for 1995 in table A1 and in table 2 are likely to be understated.

³ These two divisions together account for only about 6% of manufacturing exports in 1995, and about 4% in 2000.

⁴ The data for 2001-2004 in table A1 suggest that this omission might bias the water pollution intensity of ISIC 36 upwards, but its impact on air pollution intensity is unclear.

TableA1. Pollution Intensity of Chinese Industrial Output, 1995 and 2004 by Industry (Chinese classification GB/T 4754-2002)

Division	1995				2004			
	COD	SO2	Smoke	Dust	COD	SO2	Smoke	Dust
	<i>(kilos per thousand yuan output, 1995 constant yuan)</i>				<i>(kilos per thousand yuan output, 1995 constant yuan)</i>			
13 Agricultural and Sideline Foods Processing	13.30	2.43	2.29	0.23	1.87	0.55	0.77	0.06